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5	Mechanistic study of fulvic acid assisted photodegradation of propranolol in
6	aqueous solution
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14	Abstract
15	Flash photolysis (355 nm) and steady-state photolysis (365 nm) were applied to study mechanism
16	of propronalol photolysis in the presence of fulvic acid in aqueous solution. FA-assisted
17	photodegradation of propranolol was observed under the irradiation in UV-A region in which
18	propranolol itself is stable. Direct evidence was obtained that this photodegradation is due to static
19	quenching of FA triplet state by propranolol via electron transfer mechanism. The triplet state yield
20	($\phi_T \approx 0.6$ %) and the T-T absorption coefficient ($\epsilon_T^{620nm} \approx 5 \times 10^4 \text{ M}^{-1} \text{cm}^{-1}$) were estimated for the
21	first time from modeling of yields of FA triplet state in the presence of propranolol. Thus, fulvic
22	acid could be a promising agent for acceleration of propranolol photodegradation in aqueous
23	solution under UV-A light irradiation.
24	Keywords:

Propranolol, fulvic acid, aqueous solution, laser flash photolysis, triplet state,
 photodegradation

27 **1. Introduction**

Fulvic (FA) and humic (HA) acids are common products of natural organic matter biodegradation 28 29 and represent the main organic fraction of natural waters [1]. It is generally assumed that HA and 30 FA can transfer light energy to other components of water solutions strongly affecting the 31 photolysis of dissolved pollutants [2-5]. It is usually assumed that excited triplet states [6] and 32 active oxygen species (AOS) [7, 8] are the main active intermediates formed by UV-light excitation 33 of HA and FA. These intermediates can significantly accelerate the photodegradation of organic 34 pollutants dissolved in natural waters [9, 10]. In general FA are more effective photosensibilizators than HA [11], thereby FA could be promising agent for photochemical water purification. 35

36 Propranolol ((RS)-1-(1-methylethylamino)-3-(1-naphthyloxy)propan-2-ol, further PR) is a beta-37 blocker widely used in cardiovascular disease treatment [12]. It cannot be totally eliminated in 38 sewage treatment plants and therefore is widely detected in surface waters [13, 14]. It is shown to 39 be toxic to fish, crustaceans and green algae [15-18], which proves the ecological importance of the 40 investigation of PR transformation in environment (including photodegradation). Previous studies 41 have shown that under the sunlight irradiation PR half-life time is 0.6 to 1.9 days in summer and 3.9 42 to 12 days in winter [19], its degradation products are less toxic [20] and loose beta-adrenergic receptor specificity [21]. Direct photolysis of PR is effective only at UV-C and UV-B irradiation (λ 43 44 < 315 nm) and leads to formation of PR triplet state and hydrated electron - carbon radical cation 45 pair [22]. Observed outcomes of direct photolysis include different products of PR hydroxylation 46 [23], ring oxidation, rearrangement and deoxygenation [19].

In natural conditions indirect photodegradation of PR should prevail on direct photolysis. It was shown that PR is readily oxidized by OH radicals produced by nitrate photolysis [11] and that iron ions and hydrogen peroxide accelerate photodegradation of the related beta-blockers atenolol and metoprolol in photo-Fenton reactions [24]. Recently it was shown [11] that PR degradation is also enhanced by the presence of humic substances. Authors proposed mechanism of such enhancement: electron transfer from N-atom of amino group of PR to excited triplet state of FA (³FA, probably ketone-like) followed by a rapid hydrogen transfer to form a carbon-centered radical intermediate. Then, C-O bond is cleaved and naphthalene-1-ol is formed as the main product. It is worth to note that this mechanism was inferred based mainly on results of final product analysis and influence of oxygen content on PR photodegradation rate. We could not find any published direct evidence of ³FA formation and registration of its reaction with PR by time-resolved technique.

In our current study we aim to prove the proposed mechanism of FA assisted photodegradation of PR using nanosecond flash photolysis and to estimate the efficiency of photochemical processes involving humic substances in natural waters. We paid special attention to distinguish between two main possibilities: either ³FA directly reacts with PR or triplet state interacts with dissolved oxygen and forms AOS, which then oxidizes the drug.

63 **2. Materials and methods**

FA (C, 51.04%; H, 5.22%; O, 39.13%; N, 4.62% [25]) (Henan ChangSheng Corporation),
propranolol (hydrochloride salt) (99%, Acros) were used without additional purification.

PR concentration in working solutions was varied from 2×10^{-6} M to 2×10^{-4} M: proper volumes of 66 67 stock PR solution and stock FA solution were mixed and water was added to the desirable volume. 68 FA concentration in working solutions was 15 mg/L (for steady-state photolysis) or 30 mg/L (for 69 laser flash photolysis). Accumulation of 1,4-naphthoquinone (a main final product of PR 70 photodegradation) was monitored by appearance of its characteristic absorption band at 250 nm in 71 differential absorption spectra of photolyzed solutions. Absorption spectra of PR (there is no significant absorbance at $\lambda > 330$ nm), FA and 1,4-naphthoquinone in aqueous solution are shown 72 73 in Fig. 1.

Concentrated solutions of chemically pure sodium hydroxide were used for pH adjustments (natural
pH of FA solutions was equal to 4.5). pH measurements were carried out at ANION-4100 pH-meter
(Smolensk, Russia) with combined electrode ESK-10614.

UV-spectra were recorded on Agilent 8453 spectrophotometer ("Agilent Technologies"). Stationary photolysis was performed using high-pressure mercury lamp (DRSh-500) with water and glass filters for 365 nm mercury line separation. Quantum yield of PR photo-induced degradation was estimated from absorption of 1,4-naphthoquinone at 250 nm taking into account the known difference of absorption coefficient of PR (ϵ^{250nm}_{PR}) and 1,4-naphthoquinone (ϵ^{250nm}_{NQ}): $\epsilon^{250nm}_{NQ} - \epsilon^{250nm}_{PR} = 21300 \text{ M}^{-1} \text{cm}^{-1}$.

A set-up for laser flash photolysis (LFP) which utilizes third harmonic (355 nm) of a Nd:YAG laser (LS-2137U, Solar, Belarus) as an excitation source (pulse duration 6 ns, pulse energy 1-20 mJ, time resolution ca. 50 ns) [26] was used in the experiments. Lamp and laser intensities were measured with SOLO 2 laser power meter (Gentec EO). Absorption of ³FA in LFP experiments was monitored at characteristic wavelength 620 nm in differential transient absorption spectra.

88 Unless otherwise stated, all photochemical experiments were performed in a 1 cm quartz cell at 298
89 K in air- or argon-saturated aqueous solutions.

90 **3. Results and discussions**

91 **3.1. Laser flash photolysis of FA in the presence of PR**

We observed that excitation of FA leads to the triplet state formation ($\lambda_{max} = 620$ nm) with yield maximum in circumneutral solutions [27]. Kinetics of ³FA decay in argon-saturated solutions has a complex nature and could be adequately fitted by a set of three exponents with characteristic time constants of 4.3, 54 and 830 µs.

The excitation of FA solutions at pH 7 in the presence of PR decreases initial amplitude of FA T-T absorption at 620 nm (ΔA^{620nm}), which occurs due to the static ³FA quenching by this amine (Fig. 2). At PR concentration $\geq 2 \times 10^{-4}$ M ΔA^{620nm} reaches a stationary value. Thus, there is a fraction of ³FA, which does not react with PR or is not accessible to PR quenching. It is worth noting that ΔA^{620nm} in air-saturated solution is slightly lower than in deoxygenated one, which implies an additional channel of ³FA static quenching with dissolved oxygen (Fig. 2).

102 Figure 3 exhibits evolving intermediate absorption spectra observed at excitation of argon-saturated FA solution in the presence of 2×10^{-4} M PR. Immediately after excitation, we observe a broad 103 104 unstructured absorption band with maximum at 670 nm, which corresponds to FA T-T absorption 105 [27]. T-T absorption maximum shifts and spectrum shape changes presumably resulting from 106 selective quenching of FA triplet states (absorbing at 340-650 nm) by the complexed PR. After that 107 T-T absorption intensity decreases, absorption maximum shifts slightly towards shorter 108 wavelengths and the bleaching in UV part of spectrum decays due to partial recovery of FA ground state. It worth to note that addition of PR weakly affects the kinetics of ³FA decay (Fig. 3b), which 109 110 indicates that rate constant of dynamic quenching is low. Unfortunately, our LFP data are not 111 definitely conclusive about the role of T-T energy and electron transfer processes (reactions 1 and 2 respectively) in ³FA quenching by PR as absorption bands of cation-radical (max at 380 nm) and 112 113 triplet state (max at 410 nm) [22] of PR are not registered on the background of FA ground state bleaching at 340-400 nm. It seems that the main channel of ³FA decay is a physical quenching by 114 115 PR (reaction 3) and reactions (1) and (2) are not effective. To evaluate their contribution we performed stationary photolysis of the system FA + PR. Below we discuss these experiments in 116 117 detail.

$$118 \quad {}^{3}FA + PR = {}^{1}FA + {}^{3}PR \tag{1}$$

$$119 \quad {}^{3}FA + PR = FA^{\bullet} + PR^{+\bullet}$$
(2)

$$120 {}^{3}FA + PR = {}^{1}FA + PR$$
(3)

121 **3.2.** Stationary photolysis (365 nm) of FA aqueous solutions in the presence of PR

Fig. 4b shows the evolution of differential absorption spectra in the FA + PR system (air-saturated solution, dissolved oxygen concentration is 2.8×10^{-4} M [28]) observed during stationary photolysis. Absorption bands corresponding to PR (maxima at 210 and 290 nm) disappear and new band appears at about 250 nm indicating the accumulation of 1,4-naphthoquinone (Fig. 1) as a main final photoproduct [22].

We performed experiments with different initial PR concentrations: 4×10^{-5} M (60% of static 127 quenching) and 2×10^{-4} M (80% of static quenching) in air-saturated solutions. The results indicate 128 that with increase of amine concentration from 4×10^{-5} M to 2×10^{-4} M formation rate of final 129 130 photoproduct absorption at 250 nm increases from 0.01 per hour to 0.032 per hour (Fig. 5). Assuming that 1,4-naphthoquinone is a main photodegradation product one can estimate the 131 quantum yield of PR photo-induced degradation as 1.1×10^{-5} and 3.5×10^{-5} , accordingly. These 132 values are in fair agreement with ones published in [11]. Correlation of PR degradation quantum 133 vield with amine concentration allowed us to propose that the main channel of PR photodegradation 134 is connected with direct electron transfer from N-atom of amino group in complexed PR to excited 135 136 triplet state of FA [11] rather than with AOS generation.

To support this suggestion we performed stationary photolysis of FA + PR system in argon – saturated solution (Fig. 4a). Observed spectral changes indicate that presence of oxygen is not necessary for photo-induced PR degradation. However, absorption spectra and nature of final photoproducts in argon-saturated solution differ from ones in air-saturated solution (Fig. 4a, b). Thus, dissolved oxygen takes part in reactions with active intermediates of PR degradation or in oxidation of primary photochemical products. Detailed study of this question is beyond of the scope of current article.

144 **3.3. Modeling of triplet state quenching process with PR**

Let us suggest that ³FA static quenching by PR is connected with FA-PR complexation. In the first approximation, we can assume that all binding constants are equal and each molecule of FA can form a number of independent 1:1 complexes with PR. In this approximation we can express concentrations of occupied (Complex) and free (FA) binding sites as:

149
$$\frac{[\text{Complex}]}{[\text{FA}][\text{Prop}]} = K_{\text{st}} = \frac{[X]}{[C_{\text{FA}} - X][C_{\text{Prop}} - X]}$$
(4)

150 where C_{FA} is effective concentration of PR binding sites. X can be found by formula (5):

151
$$[X] = \frac{(C_{FA} + C_{Prop} + K_{st}^{-1}) - \sqrt{(C_{FA} + C_{Prop} + K_{st}^{-1})^2 - 4C_{FA}C_{Prop}}}{2}$$
(5)

152 We may disunite PR binding sites and triplet state formation sites and differ their total number. 153 However, for the sake of simplicity, we will further assume that one complex formed can quench β FA triplet states. Parameter β accounts for the fact that one binding center can quench several ³FA 154 as well as for the presence of ³FA not accessible for guenching by PR due to energy and/or steric 155 156 reasons. As the absorption coefficient at excitation wavelength (365 nm) is equal for complexed and 157 free FA, the amount of absorbed light does not change on varying PR concentration. Thus, changes 158 of ³FA yield are connected only with the process of static quenching of these states by complexed 159 PR. The amount of quenched triplet centers will be equal to βX , and concentration (C_T(X)) and the absorption at 620 nm ($\Delta A(X)$) of ³FA escaped quenching will be respectively: 160

161
$$\mathbf{C}_{\mathsf{T}}(\mathbf{x}) = \mathbf{C}_{\mathsf{T}} - \beta \mathbf{X}; \quad \Delta \mathbf{A}(\mathbf{X}) = \Delta \mathbf{A}_{\mathsf{0}} - \varepsilon_{\mathsf{T}}^{620\mathsf{nm}} \mathsf{I}\beta \mathbf{X}$$
 (6)

where ΔA_0 – initial amplitude of T-T absorption in the absence of PR, ε_T^{620nm} – coefficient of T-T absorption at 620 nm, 1 – optical path length. Using formulae (5) and (6) we get the final expression (7):

$$\frac{\Delta A_{0}}{\Delta A(X)} = \frac{\Delta A_{0}}{\Delta A_{0} - \varepsilon_{T}^{620nm} I\beta X} = \frac{2}{2 - \gamma [(1 + \frac{C_{Prop}}{C_{FA}} + \frac{K_{st}^{-1}}{C_{FA}}) - \sqrt{(1 + \frac{C_{Prop}}{C_{FA}} + \frac{K_{st}^{-1}}{C_{FA}})^{2} - 4\frac{C_{Prop}}{C_{FA}}]}$$
(7)

Thus, $\Delta A_0/\Delta A(X)$ is determined by three parameters – ratio of PR and FA binding center concentrations, reverse product of stability constant to binding center concentrations and $\gamma = \epsilon_T^{620nm} |C_{FA}\beta/\Delta A_0|$. The latter parameter accounts for relative decrease of T-T absorption yield when all binding centers are occupied. Inset for Fig. 2 shows a smooth curve representing the best fit of data for argon-saturated solutions according to formula (7) with parameters $\gamma = 0.83$, C_{FA} $= 10^{-6}$ M и K_{st} = 1.5×10^5 M⁻¹. Same parameters fit well the dependence of T-T absorption amplitude on PR concentration (formula (6), Fig. 2). So one can assume that this simple model is adequateenough for description of observed experimental data.

Knowing γ , the initial amplitude of T-T absorption signal, binding sites concentration and assuming $\beta = 1$ one can estimate effective coefficient of T-T absorption, which is equal to $\varepsilon_T^{620nm} \approx 5 \times 10^4 \text{ M}^ ^{1}\text{cm}^{-1}$. The latter value and the product $\phi_T \varepsilon_T^{620nm} \approx 3.3 \times 10^2 \text{ M}^{-1}\text{cm}^{-1}$, determined in [28], allows one to estimate the ³FA quantum yield as $\phi_T = 0.6$ %. Comparing the quantum yield of PR degradation (3.5×10^{-5}) and ϕ_T one can conclude that less then 1% of ³FA takes part in processes (1) and (2) leading to PR degradation. This finding can explain why it is impossible to detect cation-radical or triplet of PR in LFP experiments.

We should note that literature data on molar absorption coefficients and quantum yields for triplet states of humic substances is scarce. There are only few papers where quantum yield values were estimated as 0.3-1% depending on excitation wavelength [29]. This estimate is very important to understand the efficiency of photochemical processes involving humic substances in natural waters.

185 **Conclusions**

186 We performed mechanistic study of propronalol photodegradation in the presence of fulvic acid in 187 aqueous solution using both flash (355 nm) and steady-state photolysis (365 nm) techniques. We obtained direct evidence that FA-assisted photodegradation is due to static quenching of ³FA by 188 189 propranolol via electron transfer mechanism. We found that less than 1% of ³FA total quantity takes part in PR degradation and the quantum yield of this process is 3.5×10^{-5} . For the first time we 190 obtained the estimates of the ³FA quantum yield ($\phi_T \approx 0.6$ %) and the T-T absorption coefficient 191 $(\epsilon_T^{620nm} \approx 5 \times 10^4 \text{ M}^{-1} \text{ cm}^{-1})$ from modeling of yields of ³FA in the presence of propranolol. Humic 192 193 substances could be important factor for acceleration of propranolol photodegradation in natural 194 waters under UV-A light irradiation.

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275 Figures

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- Fig. 1. Ultraviolet/visible absorption spectra of PR (C = 1×10^{-4} M) (1), FA (C = 15 mg/L) (2) and 1,4-naphthoquinone (C = 4×10^{-5} M) (3); structural formulae of PR (1) and 1,4-naphthoquinone (3).
- 278 **Fig. 2.** The dependence of the T-T absorption yield at 620 nm on PR concentration in argon-

(circles) and in air-saturated (points) solutions. FA concentration is 30 mg/L, pH 7. Inset - the ratio

- of amplitudes of T-T absorption signal at a zero and current PR concentration depending on [PR] in
- argon- (circles) and in air-saturated (points) solutions. Smooth curves are modeling by formulas (5)
- and (6) with the parameters specified in paragraph 3.3.
- **Fig. 3.** Laser flash photolysis of FA (30 mg/L) with PR (2×10^{-4} M) in argon-saturated aqueous
- solution at pH 7. (a) Transient absorption spectra recorded at 0.4 (1), 1.2 (2), 4 (3) and 48 (4) µs
- 285 after excitation. (b) Kinetic curves at 380 (1), 520 (2) and 660 (3) nm.
- Fig. 4. Differential optical spectra, registered at 1 (1), 2 (2), 4 (3), 7 (4) and 9 (5) hours of irradiation at 365 nm (8.1 mW) during stationary photolysis of FA (C = 15 mg/L) with PR (4×10^{-5} M) in argon- (a) and in air-saturated (b) solutions at pH 7.
- Fig. 5. The dependence of optical density at 250 nm during stationary photolysis of FA (15 mg/L) with different initial PR concentrations: (circles) 4×10^{-5} M and (squares) 2×10^{-4} M in air-saturated solutions at pH 7.
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Fig. 5.